



Impact of starch-based bioplastic on growth and biochemical parameters of basil plants

This is the peer reviewed version of the following article:

Original:

Celletti, S., Fedeli, R., Ghorbani, M., Loppi, S. (2023). Impact of starch-based bioplastic on growth and biochemical parameters of basil plants. SCIENCE OF THE TOTAL ENVIRONMENT, 856(2) [10.1016/j.scitotenv.2022.159163].

Availability:

This version is available <http://hdl.handle.net/11365/1217394> since 2022-10-07T16:35:02Z

Published:

DOI: <http://doi.org/10.1016/j.scitotenv.2022.159163>

Terms of use:

Open Access

The terms and conditions for the reuse of this version of the manuscript are specified in the publishing policy. Works made available under a Creative Commons license can be used according to the terms and conditions of said license.

For all terms of use and more information see the publisher's website.

(Article begins on next page)

1 **Impact of starch-based bioplastic on growth and biochemical parameters of basil plants**

2

3 Silvia Celletti ^{a,*}, Riccardo Fedeli ^a, Majid Ghorbani ^a, Stefano Loppi ^{a,b}

4 ^a *Department of Life Sciences (DSV), University of Siena, 53100 Siena, Italy*

5 ^b *BAT Center – Interuniversity Center for Studies on Bioinspired Agro-Environmental Technology,*
6 *University of Naples “Federico II”, 80138 Napoli, Italy.*

7

8 **Abstract**

9 The recent use of bioplastics in agriculture is considered an ecological choice, aimed at limiting the
10 environmental impact of plastics, in line with the Sustainable Development Goals of the United
11 Nations. However, the impact of bioplastic residues on the environment is unclear as knowledge is
12 lacking. This is the first study investigating the effect of a starch-based bioplastic on the growth and
13 biochemical parameters of basil. Bioplastic was experimentally prepared and added to the soil at
14 2.5% (w/w), corresponding to twice the concentration of plastic mulch film residues currently found
15 in cultivated soils, in view of the increasing agricultural use of bioplastics. Basil plants were grown
16 without (controls) and with bioplastic addition for 35 days, under controlled experimental
17 conditions. Compared to the control, plants exposed to bioplastic showed stunted growth (in terms
18 of shoot fresh weight, height, and number of leaves). Significant reductions in the content of
19 chlorophyll, protein, ascorbic acid, and glucose were also observed. Finally, the treatment caused
20 oxidative stress, as evidenced by the increased content of malondialdehyde in the shoots. The
21 addition of bioplastic increased the electrical conductivity and reduced the cation exchange capacity
22 of the cultivation soil. These results suggest that bioplastic in soil may promote the onset of
23 stressful conditions for plant growth in a similar manner to plastic. They will be complemented by
24 further investigations to unravel the mechanisms underlying these responses, involving different
25 doses and types of bioplastics and other crop species.

26

27 **Keywords:** antioxidant; bio-based plastic; biodegradable plastic; corn starch; lipid peroxidation.

28

29 **1. Introduction**

30 Plastic pollution is one of the most serious and pressing environmental concerns, as plastic is an
31 indispensable resource from which affordable and useful products are obtained to satisfy the needs
32 of human society (Andrady and Neal, 2009). However, it is also an emblem of waste, pollution and
33 ecotoxicity, being an artificial polymeric material derived from fossil fuels such as petroleum (a
34 source that will run out) and not readily biodegradable (Amobonye et al., 2021; Thompson et al.,
35 2009). Furthermore, plastic has been recognized as hazardous to natural as well as agricultural
36 ecosystems and human health (Hartmann et al., 2019; Ullah et al., 2021), as it is pervasive,
37 ubiquitous and accumulates (Lebreton et al., 2018), by miniaturising into tiny particles, known as
38 microplastics (MPs) and nanoplastics (NPs) (Thompson et al., 2004).

39 Due to the growing awareness of the importance of environmental sustainability and ecological
40 transition, in 2015, all UN Member States negotiated the “2030 Agenda for Sustainable
41 Development”, to promote human well-being and protect the environment, with a focus on reducing
42 the carbon footprint and dependence on fossil fuels, in a broader effort to mitigate climate change.
43 In this context, extensive research has been devoted to exploring industrial techniques to produce
44 “green” materials, which are not harmful to the environment but have the same favourable
45 characteristics as plastic (Moshood et al., 2022). Among these, bioplastic has attracted considerable
46 attention as it is a type of polymeric material that is either biodegradable or bio-based (made at least
47 partly from biological matter such as renewable feedstocks, *e.g.*, agricultural biomass) or has both
48 characteristics. Most bioplastics currently produced belong to the group of 100% biodegradable and
49 renewable feedstocks (as for instance the starch blends), having two main advantages compared
50 with conventional plastics: (i) they decompose much faster (in 4-5 years on average, depending on
51 the chemical composition) and are therefore also easier to recycle, requiring lower energy costs; (ii)

52 being derived from biomass waste, they do not present the problem of feedstock depletion from a
53 circular economy perspective (Lamberti et al., 2020; Rosenboom et al., 2022).

54 Bioplastics are attractive in packaging for the food sector (not only for environmental protection,
55 but also for food safety) and in numerous medical and biomedical applications (Parisi et al., 2015).
56 Worthy of attention is the use of bioplastic in agriculture (Coppola et al., 2021). A potentially
57 important source of bioplastic to cultivated soils are the mulch films, which are now largely
58 manufactured from starch-based bioplastics and their use plays undoubtedly a valuable role in
59 reducing residual plastic pollution in agricultural soils and thus significantly mitigating the impact
60 that plastic residues have on crop quality and yield (Colzi et al., 2022).

61 It has been suggested that, like conventional fossil-based plastics, also bioplastic may be of
62 environmental and health concern owing to the release and decomposition of substances (*i.e.*,
63 additives and toxic chemicals) into small molecules, such as monomers and oligomers (Spaccini et
64 al., 2016; Zimmermann et al., 2020). Furthermore, bioplastics could disintegrate even faster than
65 traditional plastics and could adsorb many pollutants with various physico-chemical effects, thus
66 representing an additional threat (Wang et al., 2022). Although bioplastics can change soil
67 properties, affect crop growth and yield (Jiang et al., 2017; Zhang et al., 2016), and potentially enter
68 the food chain (Huerta Lwanga et al., 2017), to the best of our knowledge, only a very few
69 experimental studies (Huerta-Lwanga et al., 2021; Liwarska-Bizukojc, 2021; Meng et al., 2021;
70 Mroczkowska et al., 2021; Qi et al., 2018; Rillig et al., 2019; Sforzini et al., 2016; Wang et al.,
71 2022) have so far evaluated the effects of bioplastics on agroecosystems, also reporting quite
72 controversial results, sometimes showing stimulating effects (Abe et al., 2022; Huerta-Lwanga et
73 al., 2021; Liwarska-Bizukojc, 2021; Mroczkowska et al., 2021) and, in other cases, inhibiting and
74 toxic effects (Abe et al., 2022; Huerta-Lwanga et al., 2021; Liwarska-Bizukojc, 2021; Meng et al.,
75 2021; Qi et al., 2018; Wang et al., 2022). Therefore, based on the limited evidence of effects,
76 especially of starch-based bioplastic, this is a completely new scenario that needs urgently to be

77 investigated to shed light especially on the effects of bioplastics in the plant-soil system before the
78 use of these materials becomes excessive and can cause ecotoxicity.
79 In this regard, the aim of this study was to investigate whether the addition of a corn starch-based
80 bioplastic to the soil affected the growth and biochemical parameters of basil (*Ocimum basilicum*
81 L.), which was chosen as model crop species, being a very important medicinal plant and culinary
82 spice, widely cultivated in many countries under natural and greenhouse conditions and marketed
83 fresh, dried, or frozen. In view of the increasingly massive use of bioplastics in agriculture, soil was
84 therefore supplemented with approximately double the concentration of bioplastic (2.5%, w/w)
85 compared to formulations with the highest concentration described in the literature (on average
86 about 1.4%, w/w) to simulate plastic mulch film residues in cultivated soils (Meng et al., 2021; Ng
87 et al., 2018; Qi et al., 2018; Sforzini et al., 2016). Analyses of changes in soil characteristics caused
88 by bioplastic were accompanied by analyses performed on a series of useful physiological and
89 biochemical indicators related to plant growth and health status.

90

91 **2. Materials and methods**

92 *2.1. Bioplastic preparation*

93 The biodegradable corn starch-based bioplastic was made in our laboratory by the casting technique
94 according to similar methodologies (de Azevedo et al., 2020; Shafqat et al., 2021), with some
95 modifications. Briefly, 3.75 g of corn starch powder (practical grade, Saint Louis, MO, USA) were
96 mixed vigorously in 25 mL of distilled water until a homogenous white dispersion was formed.
97 Subsequently, 3.75 g of glycerol ($\geq 99.5\%$, Honeywell, Muskegon, MI, USA) and 2.5 mL of glacial
98 acetic acid (100%, Merk KGaA, Darmstadt, Germany) were added to the mixture; in particular, the
99 former increases the flexibility of the bioplastic as it acts as a plasticiser, while the latter dissolves
100 the starch more easily as it adds ions to the mixture. The resulting milky suspension was heated to
101 85 °C until it thickened and became transparent and clearer. At this step, gentle agitation was
102 carried out continuously with a glass rod to avoid the formation of bubbles and lumps. This soft

103 gelled paste was poured while still hot and immediately spread onto a glass plate (diameter = 15
104 cm) to obtain a bioplastic film (Fig. 1). Before manually peeling off, the film was first incubated to
105 harden in an oven at 100 °C for 1 h and then left to dry completely at room temperature (25 °C) for
106 one week before use.

107

108 *2.2. Potting soil preparation*

109 The soil utilized in this study was a commercial growing medium containing different components
110 (*i.e.*, acid peat, compost-free soil organic conditioner, pumice, perlite, organic fertilizer), purchased
111 from VigorPlant Italia srl, and with the following physical-chemical characteristics: 43% of
112 moisture content; 92% of porosity; 5.30 ± 0.03 of $\text{pH}_{(\text{H}_2\text{O}:1:20, \text{w/v})}$; 1.12 ± 0.01 mS cm^{-1} of electrical
113 conductivity [$\text{EC}_{(\text{H}_2\text{O}:1:20, \text{w/v})}$]; 56.89 ± 2.67 $\text{meq } 100 \text{ g}_{\text{DW}}^{-1}$ of cation exchange capacity (CEC).

114 Glass pots (diameter = 5 cm, height = 7 cm) were covered with an aluminum foil and filled with 80
115 g of soil without (C = control) or supplemented with 2.5% (w/w) bioplastic (B = bioplastic),
116 previously cut into pieces of uniform size (approximately 5 mm) using sharp blades and scissors.

117 In view of the increasingly massive use of bioplastics in agriculture, B soil was supplemented with
118 about twice the concentration of bioplastic described in the literature (on average about 1.4%, w/w),
119 to simulate plastic mulch film residues in cultivated soils (Meng et al., 2021; Ng et al., 2018; Qi et
120 al., 2018; Sforzini et al., 2016).

121 Pots were initially irrigated to 60% water holding capacity. During the plant growth period, this
122 condition was maintained by weighing each pot daily and by adding water when necessary.

123

124 *2.3. Plant growth*

125 Basil plants (cv. Riviera Ligure) were obtained by seeds previously soaked in distilled water for 1 h
126 and then germinated in darkness at 22 °C between layers of distilled water-soaked paper.
127 Subsequently, homogeneous 4-day-old seedlings were transplanted into the C and B-treated pots (3
128 seedlings/pot) and after 11 days from transplanting only one seedling per pot was left. Plants were

129 grown for 35 days (which is the typical growth stage for farm supplies in the Italian market)
130 (Sgherri et al., 2010) under controlled experimental conditions [temperature (25/20 °C, day/night),
131 relative humidity (70%), photoperiod (16/8 h, day/night), and light intensity (250 $\mu\text{mol m}^{-2} \text{s}^{-1}$
132 PAR)] in a climatic chamber. To account for possible microclimatic conditions within the growth
133 chamber, the pots were randomly rotated every day.

134

135 *2.4. Plant analyses*

136 *2.4.1. Chlorophyll and biometric parameters*

137 At the end of plant growth period (corresponding to harvest: 35 days from seedling transplanting),
138 the chlorophyll content of basil leaves was measured using a portable and non-destructive
139 chlorophyll content meter (CCM – 300, Opti-Sciences Inc, Hudson, NH, USA). Specifically, 6
140 values per leaf were measured at pairs of points diametrically opposite the midrib in the following
141 order: two along the proximal, two along the central, and two along the distal part from the base to
142 the apex of the leaf, thus 12 measurements per plant were taken on the youngest and fully expanded
143 leaves. The chlorophyll content was expressed on a surface basis (mg m^{-2}).

144 Before harvest, several biometric parameters were recorded. The plant height was measured with a
145 ruler (considering the distance between the plant apex and the soil surface), the number of leaves
146 was counted, and the aboveground biomass was first weighted (expressed in terms of shoot fresh
147 weight) and then immediately frozen at 20 °C for subsequent analyses.

148

149 *2.4.2. Proteins and sugars*

150 The contents of total soluble proteins and soluble sugars (fructose, glucose, and sucrose) were
151 determined in the extracts of basil shoots obtained following the methods described in Fedeli et al.
152 (2022), with slight modifications. Specifically, a total of 0.250 g of frozen material was
153 homogenised in 1.5 mL of distilled water. The extract solution was centrifuged (PK110 centrifuge,
154 Alc International S.r.l., Cologno Monzese, MI, Italy) at 3000 rpm for 5 min at room temperature.

155 The supernatant was recovered and centrifuged again at 12000 rpm (Z 233 MK-2, Hermle,
156 LaborTechnik GmbH, Wehingen, Germany) for 7 min at room temperature.

157 For protein determination, an aliquot of the extract (20 μ L) was diluted to 1 mL with distilled water
158 and then 0.4 mL of the diluted sample were combined with 1.6 mL of the Bradford dye reagent
159 solution (Thermo Fisher Scientific Inc., Waltham, MA, USA) (Bradford, 1976). After 20 min, the
160 absorbance of the samples was measured at 595 nm using a UV-Vis spectrophotometer (8453,
161 Agilent, Santa Clara, CA, USA). Results were expressed as mg g^{-1} using bovine serum albumin
162 (BSA) (Sigma-Aldrich, USA) as standard.

163 For sugar determination, the remaining extract was filtrated using 45 μ m syringe filters (diameter =
164 25 mm, Lab Logistic Group GmbH, Meckenheim, Germany) and 150 μ L of the filtrate was
165 transferred in new tubes placed in a vacuum evaporator (Jouan RC 10-10 Vacuum Concentrator
166 Centrifugal Evaporator, Analytical Instruments Brokers LLC, Golden Valley, MN, USA) at 40 $^{\circ}$ C
167 until completely dried. Subsequently, the samples were resuspended in 30 μ L of distilled water and
168 directly analysed by HPLC (600E System, Waters, Milford, MA, USA). Sugar separation was
169 allowed using distilled water as mobile phase at a flow rate of 0.5 mL min^{-1} and an ion-exchange
170 column (10 μ m, 300 \times 6.5 mm, Sugar-Pak I, Waters, Milford, MA, USA), kept constantly at 90 $^{\circ}$ C
171 by means of an external temperature controller (Column Heater Module, Waters, Milford, MA,
172 USA). The sugars were detected by a refractive index detector (2410 RI, Waters, Milford, MA,
173 USA). Quantification was obtained by preparing individual stock solutions, using sugar reagent-
174 grade analytical standards (D-Fructose, α -D-Glucose, Sucrose, Merk KGaA, Darmstadt, Germany).

175

176 2.4.3. *Ascorbic acid*

177 The ascorbic acid (vitamin C) content was estimated colorimetrically following the method by
178 Jagota and Dani (1982), with some modifications. In brief, 0.8 mL of 10% (w/v) trichloroacetic acid
179 (TCA) (99.5%, Panreac, Castellar del Vallès, Barcellona, Spain) extraction solution were added to
180 0.2 g of frozen shoot material. The samples were homogenised with an ULTRA-TURRAX[®] (T 10

181 basic, Werke GmbH & Co. KG, Staufen, Germany) prior to filtration on gauze. Subsequently, the
182 filtrates were kept in an ice bath for 5 min and then centrifuged (PK110 centrifuge, Alc
183 International S.r.l., Cologno Monzese, MI, Italy) at 3000 rpm for further 5 min at room temperature.
184 An aliquot of 0.4 mL of supernatant was transferred into tubes containing 1.6 mL of distilled water.
185 Then, 0.2 mL of 0.2 M Folin – Ciocalteu reagent (Carlo Erba, Cornaredo, MI, Italy) were added to
186 the diluted supernatants and shaken vigorously. The mixtures were incubated for 10 min at room
187 temperature and thereafter the absorbances were measured at 760 nm (8453, UV – Vis
188 Spectrophotometer, Agilent, Santa Clara, CA, USA). Calibration was done with 0.05 – 0.2 mL of a
189 100 $\mu\text{g mL}^{-1}$ L-ascorbic acid (BioXtra, $\geq 99.0\%$, crystalline) stock solution. The ascorbic acid
190 content of the samples was expressed as $\mu\text{g g}^{-1}$.

191

192 2.4.4. Malondialdehyde

193 Oxidative stress of membrane lipids was estimated by analysing the content of malondialdehyde
194 (MDA) as a metabolite reactive to 2-thiobarbituric acid (TBA) according to Quagliata et al. (2021),
195 with slight modifications. Briefly, 0.5 g of frozen shoots were homogenised with an ULTRA-
196 TURRAX[®] (T 10 basic, Werke GmbH & Co. KG, Staufen, Germany) in 5.0 mL of extraction
197 solution, obtained by dissolving 0.25% (w/v) TBA ($\geq 98.0\%$, Merk KGaA, Darmstadt, Germany) in
198 10% (w/v) TCA (99.5%, Panreac, Castellar del Vallès, Barcellona, Spain). The homogenates were
199 incubated at 95 °C for 30 min in a hot-water bath (GBath 1800 Digital Thermostatic Bath, F.lli
200 Galli G. & P., Milano, MI, Italy) and then immediately cooled on ice to stop the reaction. Once
201 completely cold, the samples were centrifuged (PK110 centrifuge, Alc International S.r.l., Cologno
202 Monzese, MI, Italy) at 5000 rpm for 20 min at room temperature. After centrifugation, absorbance
203 was measured on the recovered supernatants with a UV-Vis spectrophotometer (8453, Agilent,
204 Santa Clara, CA, USA). To correct the absorbance at 532 nm from the interference of non-specific
205 turbidity, absorbance at 600 nm was subtracted from the reading. The MDA content was expressed

206 as $\mu\text{g g}^{-1}$ using the molar extinction coefficient of the formed MDA – TBA complex of 155 mM^{-1}
207 cm^{-1} .

208

209 *2.5. Planted soil analyses*

210 After harvesting the aboveground plant biomass, the entire potting soil was taken and considered as
211 rhizosphere soil as the pots were completely rooted. Soil samples were oven-dried at $105 \text{ }^\circ\text{C}$ for a
212 week to get constant weight and then crushed to pass through a 2-mm sieve before analysis.

213

214 *2.5.1 pH and electrical conductivity*

215 The soil pH and EC were measured according to (Celletti et al., 2021b) in the limpid supernatants
216 (soil:dH₂O ratio of 1:20, g_{DW}:mL), after 5 min of centrifugation at 4000 rpm (PK110 centrifuge, Alc
217 International S.r.l., Cologno Monzese, MI, Italy) and paper filtration from a 2 h-initial shaking (711,
218 VDRL STIRREL, ASAL srl, Cernusco sul Naviglio, MI, Italy). The pH was determined using a
219 pH-meter (edge® HI2002, HANNA Instruments Inc., Woonsocket, RI, USA) and EC using a
220 conductimeter (BASIC 30, EC – meter, Crison Strumenti SpA, Carpi, MO, Italia).

221

222 *2.5.2 Cation exchange capacity*

223 The CEC was determined as described by Bascomb (1964) with some modifications. In brief, 2 g of
224 soil were weighed in centrifuge tubes and 25 mL of 10% (w/v) BaCl₂ × 2 H₂O solution buffered
225 with 8.1 pH triethanolamine solution were added and shaken for 3 min, left to rest for 5 min and
226 then shaken again for 3 min. The samples were centrifuged (PK110 centrifuge, Alc International
227 S.r.l., Cologno Monzese, MI, Italy) at 3000 rpm for 5 min at room temperature and the supernatants
228 were discarded. The sedimented soils were resuspended by adding 25 mL of distilled water, shaken,
229 centrifuged, and the supernatants discarded. Further 25 mL of 0.1 N MgSO₄ × 7 H₂O solution were
230 added to the washed soils, again shaken and centrifuged. An aliquot (10 mL) of each clear
231 supernatant was immediately transferred to a conical flask, containing 100 mL of distilled water and

232 10 mL of 30% ammonium hydroxide solution. These solutions were titrated under slow stirring
233 with 0.05 N ethylenediaminetetraacetic acid disodium salt (EDTA – Na₂) solution, using 2 drops of
234 0.1% (w/v) Eriochrome Black T indicator. In parallel, a blank sample was titrated by pipetting 10
235 mL of 0.1 N MgSO₄ × 7 H₂O solution into the flask instead of the supernatant. The endpoint of
236 sample titration was indicated by the colour change from clear blue to reddish purple. The results
237 were expressed as meq 100 g⁻¹, according to the formula:

$$238 \quad CEC\ of\ the\ soil = \frac{(m - n) * 0.05 * 2.5 * 100}{p}$$

239 where,

240 m = volume EDTA – Na₂ (mL) used to titrate the blank sample;

241 n = volume EDTA – Na₂ (mL) used to titrate the soil sample;

242 0.05 = normality of EDTA – Na₂ solution;

243 2.5 = conversion factor to relate the 10 mL titrated to the 25 mL MgSO₄ solution added;

244 p = weight of the soil sample (g).

245

246 2.6. Statistical analysis

247 Data normality was verified with the Shapiro-Wilk test. The results are presented as mean ±
248 standard error (SE) from five biological replicates (n = 5). The experiment was replicated 3 times.

249 Significant differences (p < 0.05) between C and B means were evaluated with the Student's t-test.

250 Calculations were run using the free software R version 4.0.3 (R Core Team 2022). For correlation
251 analysis, the Pearson correlation coefficient was used.

252

253 3. Results

254 After 35 days of growth in the presence of the corn starch-based bioplastic, basil plants showed a
255 significant decline (-8%) in the content of leaf chlorophyll (Fig. 2A). Bioplastic also negatively

256 affected the biometric parameters of plant growth and development: plant height, number of leaves,

257 and aboveground fresh weight were all strongly reduced by -68%, -42%, and -82% (Fig. 2B, C, and
258 D), respectively. Figure 2E displays visually these considerable differences in the growth of basil
259 plants cultivated without (C = control, on the left) and with bioplastic (B = bioplastic, on the right)
260 added in soil.

261 Only the content of glucose was significantly decreased (-22%) by the addition of bioplastic, while
262 changes were insignificant for fructose and sucrose at shoot level (Fig. 3A). The total content of
263 soluble sugars (given by the sum of fructose, glucose, and sucrose) showed an overall decreasing
264 trend after plastic supplementation.

265 The addition of bioplastic to the soil caused a significant reduction (-44%) in the content of the
266 soluble proteins in the aboveground part of basil plants (Fig. 3B).

267 Also, the content of ascorbic acid (vitamin C) in basil shoots was significantly decreased (-9%) after
268 bioplastic addition (Fig. 4A).

269 When basil plants were grown in soil added with bioplastic, the content of MDA in shoots was
270 significantly increased by 17% (Fig. 4B).

271 The correlation analysis performed between the content of ascorbic acid and MDA in the shoots
272 showed a negative linear correlation between these two biomarkers (Pearson's $r = -0.561$), but this
273 was not statistically significant ($p = 0.092$) (Fig. 4C).

274 At soil level, the presence of the corn starch-based bioplastic did not modify the pH (remaining
275 stable at ~ 7.6), even though it was more than 2 units higher than that found at the beginning of the
276 experiment (Fig. 5A); on the contrary, EC increased by 16%, while CEC decreased (-7%), albeit
277 slightly, but significantly (Fig. 5B, and C, respectively).

278

279 **4. Discussion**

280 In agriculture, the challenge of replacing conventional fossil-based plastics, commonly used to
281 produce mulch films, with bioplastics (especially bio-based and biodegradable ones, such as starch-
282 based ones) is of utmost importance for the climate, in line with the Sustainable Development Goals

283 of the United Nations 2030 Agenda. In contrast to plastics, it is not yet clear what impact the
284 massive use of bioplastics will generate on the environment, and particularly on agricultural soils,
285 both in the short- and long-term, as knowledge on these aspects is so far lacking and controversial.
286 It is known that bioplastics decompose by releasing chemicals (Spaccini et al., 2016) and, therefore,
287 like plastics, they may imbalance soil characteristics and, consequently, affect the growth and yield
288 of crops (Jiang et al., 2017; Zhang et al., 2016), eventually accumulating in the parts of edible crops
289 and entering the food chain (Huerta Lwanga et al., 2017) with possible negative implications for
290 human health (Li et al., 2020; Muncke et al., 2020). Within this scenario, aware and conscientious
291 research is therefore extremely urgent and essential to decipher the impact of bioplastics on the
292 environment according to a sustainability perspective (Lamberti et al., 2020; Rosenboom et al.,
293 2022).

294 The present study investigated the impact of a corn starch-based bioplastic, added to the soil, on the
295 growth and development of basil, which is a highly interesting food crop plant. Our study focused
296 on the analysis of the shoot alone, as this is the commercial part of interest for basil plants. The
297 bioplastic tested was obtained experimentally in our laboratory by mixing various components
298 (such as corn starch, glycerol, and acetic acid) in the appropriate proportions. Accordingly,
299 knowledge of the single components of bioplastics is an advantage which will allow to investigate
300 in the foreseeable future what component will prevail and how the different components
301 individually will affect plant growth and soil characteristics. The concentration of bioplastic was
302 established at 2.5% (w/w), based on the current use of bioplastics in agriculture (mostly in the form
303 of mulch films), which amounts to more than 1% on average (Ng et al., 2018; Qi et al., 2018;
304 Sforzini et al., 2016), but which could likely increase to extremely high levels in the near future in
305 view of a predictive increase in the use of bioplastics due to the ecological transition in this sector
306 (Meng et al., 2021).

307 The possible interaction of bioplastic with the plant was investigated by monitoring changes in
308 growth and biochemical features associated with vegetative development and plant health.

309 The reductions in chlorophyll content and fresh biomass of bioplastic-treated basil plants observed
310 in this study are consistent with the phytotoxicity effects exerted by the addition of another bio-
311 based and biodegradable bioplastic, the polylactic acid (PLA), at 10% and 2.5% (dry soil weight),
312 on the leaves of common bean and maize plants, respectively (Meng et al., 2021; Wang et al.,
313 2020). On the other hand, in the same study by Wang et al. (2020), it has been evaluated also the
314 effect of polyethylene (PE), which is a bio-based but non-biodegradable bioplastic, and it seemed
315 not to cause any phytotoxic effect on maize plants. Conversely, Pignattelli et al. (2021)
316 demonstrated that garden cress (*Lepidium sativum* L.) exposed to the lower size (61 – 499 μm) of
317 polyethylene terephthalate (PET), having the same characteristics of PE, decreased the
318 photosynthetic efficiency. Therefore, the studies mentioned above suggest that different types of
319 bioplastic polymers affected differently plant growth and that differences are most likely a function
320 of the bioplastic's level of degradability and size. In our experiments, at harvest, the added
321 bioplastic left no physical traces after careful visual inspection, indicating that it had totally
322 dissolved in the soil (Mroczkowska et al., 2021) and therefore, being readily bioavailable, was
323 probably taken up by the root system.

324 In addition, our data showing a negative regulation of plant growth parameters was in line with the
325 outcomes of recent studies (Qi et al., 2018; Sun et al., 2020; van Weert et al., 2019), reporting the
326 effects of MPs and NPs on different plant species. Hence, we can suggest that, in this case, the
327 bioplastic exerted effects similar to plastic.

328 Combining the observed negative effects on basil growth parameters by bioplastic and the fact that,
329 in plants, sugars, produced by photosynthesis, are used to support all aspects of plant growth and
330 development (Ciereszko, 2018; Sami et al., 2016), we evaluated how the content of soluble sugars
331 changed in the shoot when basil plants were grown with bioplastic. Specifically, we determined the
332 content of the disaccharide sucrose and that of both its two distinct monosaccharide constituents
333 (*i.e.*, fructose and glucose). Interestingly, only the content of glucose dropped significantly. This
334 lower glucose level could explain the lower biomass accumulation observed in the shoots of the

335 bioplastic-treated plants, as glucose acts as a signal molecule and phytohormone affecting the
336 expression of many different genes involved in key processes such as leaf growth (Moore et al.,
337 2003). Indeed, the bioplastic tested in our study seemed to behave like a type of plastic (*i.e.*,
338 polystyrene – PS) used in the experiments by S. Li et al. (2021), where PS particles, once absorbed
339 by the roots of barley plants, caused an inhibition of energy supplementation and biomass
340 accumulation. Moreover, the overall decreasing trend resulting from summing up all sugars also
341 agreed with the observations in cucumber fruits by Z. Li et al. (2021), where treatment with PSNP
342 of different sizes significantly reduced the soluble sugar content.

343 After bioplastic exposure, basil plants also reacted by reducing the content of proteins, which are
344 another class of primary organic compounds. This result is reasonable since proteins are biological
345 macromolecules constituting essential building blocks of all living organisms, including plants, and,
346 thus, a slowdown in their synthesis is a clear sign that the basil plants were in unfavourable growth
347 condition (Murray et al., 2017). To the best of our knowledge, no study in the literature has
348 documented any effect of bioplastics on changes in the protein amount in plants. Only the study by
349 Z. Li et al. (2021) demonstrated an increase of this parameter on cucumber fruits, but using a plastic
350 material, the PSNP, of different sizes. Since many proteins belong to the enzyme category, it can be
351 speculated that, under these conditions, the addition of bioplastic to the soil might have hindered
352 some biochemical enzymatic reactions of vital importance for plant metabolism, such as
353 photosynthesis, as evidenced by the significant reduction in the content of chlorophyll, an essential
354 molecule for the proper functioning of the photosynthetic process.

355 For its protective role against the effects of drought, ozone, and ultraviolet sunlight (Gallie, 2013),
356 L-ascorbic acid (commonly known as vitamin C) is the most widespread non-enzymatic antioxidant
357 compound in plants (Arrigoni and De Tullio, 2002; Ishikawa et al., 2006). Very worthy of mention,
358 Dowdle et al. (2007) demonstrated the linkage between ascorbic acid and the growth and life of
359 plants. These authors revealed ascorbic acid is biosynthesised from hexose sugars (including
360 fructose and glucose), given the discovery of the existence of a specific enzyme, GDP-L-galactose

361 phosphorylase, capable of synthesising ascorbic acid in plants (Dowdle et al., 2007). Our findings
362 suggested that lower glucose accumulation probably led to a significant decrease in ascorbic acid
363 level in bioplastic-treated basil shoots, further corroborating the evidence of the essential correlation
364 between glucose and ascorbic acid. Experiments that evaluated the effect of different PSNP sizes on
365 cucumber fruits also confirmed this relationship between the two molecules (Z. Li et al., 2021).
366 These authors clearly demonstrated that treatment with PSNP of 500 nm in size significantly
367 reduced both ascorbic acid and soluble sugar content, while treatment with PSNP of 100 nm
368 significantly increased both. In addition, these findings support our results regarding the effect of
369 bioplastic, which, like plastic, decreased both sugar and ascorbic acid content.

370 Plants subjected to environmental stress have been shown to increase the production of antioxidant
371 compounds, in order to counteract the increased production of reactive oxygen species (ROS)
372 (Hasanuzzaman et al., 2020); these species are harmful to plant vitality as they react with cell
373 membrane lipids and cause their peroxidation (Su et al., 2019). In our study, we analysed the
374 content of MDA, which is the main product formed at the end of the chain of radical reactions
375 caused by a stressful environmental condition and, therefore, it is commonly used as a biomarker
376 for detecting the extent of oxidative damage to biological membranes (Shulaev and Oliver, 2006).
377 Exposure of basil plants to corn starch-based bioplastic significantly increased the MDA content in
378 the shoots. On the other hand, however, as mentioned above, basil plants showed a rather weak
379 ascorbic acid scavenging ability to cope with the high level of oxidative stress. Indeed, at the onset
380 of oxidative stress, it has been widely observed that antioxidant compounds fail to counteract ROS
381 production, which occurs normally in plant metabolism. Certainly, a more solid hypothesis could be
382 formulated by also analysing the activities of antioxidant enzymes as well as the level of non-
383 enzymatic antioxidants, such as ascorbic acid in this case (Wani et al., 2013). Thus, this type of
384 bioplastic might have established a trade-off mechanism in the aerial part of the plant whereby
385 when oxidative stress compounds increase, defence compounds decrease, although the correlation
386 analysis between the two evaluated biomarkers did not statistically validate it. However, this

387 hypothesis is not dissimilar to the study by Pignattelli et al. (2021), describing the effects of PET
388 MP treatment in cress plants, in which an opposite trend was evident between the antioxidant
389 defence response (which decreased) and the accumulation of ROS (which increased) in the leaf
390 tissue. In contrast, Gao et al. (2019) reported that treatment with PE MP increased both the level of
391 oxidative stress and the content of ascorbic acid in lettuce plants. Based on the high levels of MDA
392 accumulation observed in the shoots and the knowledge that these high levels may damage cell
393 membranes and even lead to cell death (Sharma et al., 2012), it could be a plausible explanation that
394 the starch-based bioplastic, considered in our study, can be involved in the impairment of bio-
395 membrane proteins, as supported by the drop in total protein content found in our basil shoots.

396 In this study, we can assume that most of the effects observed at the shoot level of basil plants are
397 probably the indirect consequence of the actions exerted by bioplastic addition to the soil on the
398 proper function of plant roots. As an example, these actions may include the subtraction of oxygen
399 due to the degradation processes of the bioplastic by microorganisms and the depletion of water in
400 the soil due to the hydrophilicity and water absorption by bioplastic (Abe et al., 2021). The
401 changing physiological and biochemical responses observed in the leaf apparatus of basil plants
402 exposed to bioplastic could reflect an imbalance in some fundamental, yet crucial, soil variables.

403 Much is known about the changes in soil properties due to the persistence and accumulation of
404 plastics (Bouaicha et al., 2022; Chen et al., 2022; Khalid et al., 2020; Liu et al., 2017). Likewise
405 plastics, it should be stressed out that the assessment of the impact of bioplastic in agricultural soils
406 is of paramount importance from an ecological and human food safety perspective (Qi et al., 2018).

407 For these reasons, we analysed some inherent soil chemical characteristics, mainly pH, EC, and
408 CEC in soils where basil plants were grown. One of the most important soil variables is pH, as it
409 mainly affects the availability of nutrients to plants (Delgado and Gómez, 2016; Fageria and Stone,
410 2006). Estimating the EC of a soil can provide a lot of helpful information about the overall soil
411 health. As an example, high EC levels can mean that a soil, or even a growing medium in general,
412 contains a high content of salts (mainly sodium – Na), which are potentially harmful to plant vitality

413 (Celletti et al., 2021a; Hazelton and Murphy, 2007). On the other hand, CEC is an effective index to
414 assess soil fertility. Indeed, this parameter measures the capacity of the soil to retain exchangeable
415 positively charged ions (cations) through electrical attraction. For example, when a soil is rich in
416 organic matter, it has a very high CEC and, therefore, means that nutrients can move through the
417 soil and become available to plants (Hazelton and Murphy, 2007). In the studies by Boots et al.
418 (2019) and Yu et al. (2020), it was observed that bio-based and non-biodegradable bioplastic
419 (specifically as high-density polyethylene – HDPE and PE, respectively), altered the cation and
420 proton exchange of the soil, resulting in a drop in soil pH and CEC. On the other hand, other
421 experiments that tested PE and PLA at different dosages illustrated that these treatments increased
422 pH (Wang et al., 2020). Looking back to our results, no pH-dependent change was visible, while
423 that of CEC resulted in a significant reduction in agreement with the above-mentioned results. With
424 regard to EC, on the other hand, we noticed a considerable increase. If we combine the values of
425 increasing EC and decreasing CEC, we can generalise by hypothesising that these trends may be
426 related to an increase in salt concentration (such as monovalent cations: Na and potassium, for
427 instance) rather than to an increased release of potentially exchangeable nutrients (such as divalent
428 cations: calcium and magnesium, for instance) and, thus, absorbable by plants.

429 Thus, it was evident that the addition of a high amount of biodegradable material caused an
430 alteration of the soil environment (specifically, an increase in EC and a decrease in CEC). Taken
431 together, these changes could be related to the changes in growth and biochemical parameters
432 observed in the shoots. Indeed, the impact of high salt concentrations in the soil on the reduction of
433 plant vigour, leaf number and/or size, discoloured foliage and increased oxidative stress levels of
434 the plants is well known (Ashraf, 2009; Parida and Das, 2005; Wani et al., 2013). Furthermore,
435 excess salts (such as Na) can compete with nutrients, leading to an induction of a nutrient imbalance
436 and reducing nutrient availability to the plants (Machado and Serralheiro, 2017; Wani et al., 2013).

437 Among the nutrients, iron (Fe) plays a central role in photosynthesis and chlorophyll synthesis, as
438 well as in respiration and nitrogen assimilation (Celletti et al., 2020; Connorton et al., 2017; Rout

439 and Sahoo, 2015). Given the central role of this nutrient, the imbalance in Fe uptake due to the
440 interference of the bioplastic could be one of the main factors that triggered the reduction in plant
441 growth (Bartucca et al., 2017; Del Buono et al., 2015).

442 However, it is worth emphasizing that this is the first study investigating the impact of a corn
443 starch-based bioplastic on the growth and biochemical parameters of basil plants. To the best of our
444 knowledge, many key questions still need to be clarified, as there is currently little information
445 concerning the impact of bioplastics on plants. Therefore, this phenomenon needs to be additionally
446 explored at multiple levels to better understand the mechanisms underlying these physiological and
447 biochemical responses. Further studies will be required to identify which component is mainly
448 responsible for the effects on plants, to test the effect of this treatment on other crops and soil types,
449 also using different types of bioplastics and different concentrations, to test whether the plant
450 response is species- and/or dose-dependent, and to determine the threshold beyond which the plant
451 is strongly induced to trigger intense reprogramming of metabolic processes with high energy costs.

452

453 **5. Conclusions**

454 Looking back at the initial hypothesis, the present study verified the possible interference on basil
455 plants of the presence in the soil of residues of potentially contaminating emerging materials, such
456 as corn starch-based bioplastic, which belongs to the group of biodegradable and bio-based
457 bioplastics and is one of the most widely used on the market. The results clearly indicated that this
458 type of bioplastic affected plant growth parameters and changed the properties of planted soil.
459 Specifically, growth was stunted, the defense response weakened, and oxidative stress was induced
460 in the aerial part of basil plants. Overall, these results provided some clues as to whether bioplastic
461 added to the soil simulates the effects of plastic, favouring the onset of a stressful condition for
462 plant survival and vitality.

463 Hence, within this framework of the results presented in this study, there is a need to examine the
464 impact of bioplastics in agriculture, as there is currently little information on their effects on

465 cultivated plants. From a more general point of view, it would be forward-looking to address the
466 various challenges of evaluating bioplastic impacts to raise environmental and social awareness
467 before the use of these materials may become excessive and cause ecotoxicity problems.

468

469 **Acknowledgements**

470 We would like to thank Massimo Guarnieri for his technical help with HPLC analysis.

471

472 **Author contributions**

473 **Silvia Celletti:** Conceptualization, Data curation, Formal analysis, Investigation, Methodology,

474 Supervision, Visualization, Writing – original draft. **Riccardo Fedeli:** Data curation, Investigation.

475 **Majid Ghorbani:** Investigation. **Stefano Loppi:** Conceptualization, Resources, Supervision,

476 Writing – review & editing. All authors have read and approved the final manuscript.

477

478 **Funding**

479 This research did not receive any specific grant from funding agencies in the public, commercial, or

480 not-for-profit sectors.

481

482 **Conflicts of Interest**

483 The authors declare no conflict of interest.

484 **References**

- 485 Abe, M.M., Branciforti, M.C., Brienzo, M., 2021. Biodegradation of Hemicellulose-Cellulose-
486 Starch-Based Bioplastics and Microbial Polyesters. *Recycl.* 2021, Vol. 6, Page 22 6, 22.
487 <https://doi.org/10.3390/RECYCLING6010022>
- 488 Abe, M.M., Branciforti, M.C., Nallin Montagnolli, R., Marin Morales, M.A., Jacobus, A.P.,
489 Brienzo, M., 2022. Production and assessment of the biodegradation and ecotoxicity of xylan-
490 and starch-based bioplastics. *Chemosphere* 287, 132290.
491 <https://doi.org/10.1016/j.chemosphere.2021.132290>
- 492 Amobonye, A., Bhagwat, P., Raveendran, S., Singh, S., Pillai, S., 2021. Environmental Impacts of
493 Microplastics and Nanoplastics: A Current Overview. *Front. Microbiol.* 12, 3728.
494 <https://doi.org/10.3389/FMICB.2021.768297/BIBTEX>
- 495 Andrady, A.L., Neal, M.A., 2009. Applications and societal benefits of plastics. *Philos. Trans. R.*
496 *Soc. B Biol. Sci.* 364, 1977–1984. <https://doi.org/10.1098/rstb.2008.0304>
- 497 Arrigoni, O., De Tullio, M.C., 2002. Ascorbic acid: much more than just an antioxidant. *Biochim.*
498 *Biophys. Acta* 1569, 1–9. [https://doi.org/10.1016/S0304-4165\(01\)00235-5](https://doi.org/10.1016/S0304-4165(01)00235-5)
- 499 Ashraf, M., 2009. Biotechnological approach of improving plant salt tolerance using antioxidants as
500 markers. *Biotechnol. Adv.* 27, 84–93. <https://doi.org/10.1016/J.BIOTECHADV.2008.09.003>
- 501 Bartucca, M.L., Celletti, S., Mimmo, T., Cesco, S., Astolfi, S., Del Buono, D., 2017. Terbutylazine
502 interferes with iron nutrition in maize (*Zea mays*) plants. *Acta Physiol. Plant.* 39.
503 <https://doi.org/10.1007/s11738-017-2537-z>
- 504 Bascomb, C.L., 1964. Rapid method for the determination of cation-exchange capacity of
505 calcareous and non-calcareous soils. *J. Sci. Food Agric.* 15, 821–823.
506 <https://doi.org/10.1002/JSFA.2740151201>
- 507 Bouaicha, O., Mimmo, T., Tiziani, R., Praeg, N., Polidori, C., Lucini, L., Vigani, G., Terzano, R.,
508 Sanchez-Hernandez, J.C., Illmer, P., Cesco, S., Borruso, L., 2022. Microplastics make their
509 way into the soil and rhizosphere: A review of the ecological consequences. *Rhizosphere* 22,

510 100542. <https://doi.org/10.1016/J.RHISPH.2022.100542>

511 Bradford, M.M., 1976. A rapid and sensitive method for the quantitation of microgram quantities of
512 protein utilizing the principle of protein-dye binding. *Anal. Biochem.* 72, 248–254.
513 [https://doi.org/10.1016/0003-2697\(76\)90527-3](https://doi.org/10.1016/0003-2697(76)90527-3)

514 Celletti, S., Bergamo, A., Benedetti, V., Pecchi, M., Patuzzi, F., Basso, D., Baratieri, M., Cesco, S.,
515 Mimmo, T., 2021a. Phytotoxicity of hydrochars obtained by hydrothermal carbonization of
516 manure-based digestate. *J. Environ. Manage.* 280, 111635.
517 <https://doi.org/10.1016/j.jenvman.2020.111635>

518 Celletti, S., Lanz, M., Bergamo, A., Benedetti, V., Basso, D., Baratieri, M., Cesco, S., Mimmo, T.,
519 2021b. Evaluating the Aqueous Phase From Hydrothermal Carbonization of Cow Manure
520 Digestate as Possible Fertilizer Solution for Plant Growth. *Front. Plant Sci.* 12.
521 <https://doi.org/10.3389/fpls.2021.687434>

522 Celletti, S., Pii, Y., Valentinuzzi, F., Tiziani, R., Fontanella, M.C., Beone, G.M., Mimmo, T.,
523 Cesco, S., Astolfi, S., 2020. Physiological Responses to Fe Deficiency in Split-Root Tomato
524 Plants: Possible Roles of Auxin and Ethylene? *Agronomy* 10, 1000.
525 <https://doi.org/10.3390/agronomy10071000>

526 Chen, G., Li, Y., Liu, S., Junaid, M., Wang, J., 2022. Effects of micro(nano)plastics on higher
527 plants and the rhizosphere environment. *Sci. Total Environ.* 807, 150841.
528 <https://doi.org/10.1016/j.scitotenv.2021.150841>

529 Ciereszko, I., 2018. Regulatory roles of sugars in plant growth and development. *Acta Soc. Bot.*
530 *Pol.* 87. <https://doi.org/10.5586/asbp.3583>

531 Colzi, I., Renna, L., Bianchi, E., Castellani, M.B., Coppi, A., Pignattelli, S., Loppi, S., Gonnelli, C.,
532 2022. Impact of microplastics on growth, photosynthesis and essential elements in *Cucurbita*
533 *pepo* L. *J. Hazard. Mater.* 423, 127238. <https://doi.org/10.1016/j.jhazmat.2021.127238>

534 Connorton, J.M., Balk, J., Rodríguez-Celma, J., 2017. Iron homeostasis in plants-a brief overview.
535 *Metallomics.* <https://doi.org/10.1039/c7mt00136c>

536 Coppola, G., Gaudio, M.T., Lopresto, C.G., Calabro, V., Curcio, S., Chakraborty, S., 2021.
537 Bioplastic from Renewable Biomass: A Facile Solution for a Greener Environment. *Earth*
538 *Syst. Environ.* 2021 52 5, 231–251. <https://doi.org/10.1007/S41748-021-00208-7>

539 de Azevedo, L.C., Rovani, S., Santos, J.J., Dias, D.B., Nascimento, S.S., Oliveira, F.F., Silva,
540 L.G.A., Fungaro, D.A., 2020. Biodegradable Films Derived from Corn and Potato Starch and
541 Study of the Effect of Silicate Extracted from Sugarcane Waste Ash. *ACS Appl. Polym. Mater.*
542 2, 2160–2169. <https://doi.org/10.1021/ACSAPM.0C00124>

543 Del Buono, D., Astolfi, S., Mimmo, T., Bartucca, M.L., Celletti, S., Ciaffi, M., Cesco, S., 2015.
544 Effects of terbuthylazine on phytosiderophores release in iron deficient barley. *Environ. Exp.*
545 *Bot.* 116, 32–38. <https://doi.org/10.1016/j.envexpbot.2015.03.007>

546 Delgado, A., Gómez, J.A., 2016. The Soil. Physical, Chemical and Biological Properties, in:
547 Principles of Agronomy for Sustainable Agriculture. Springer International Publishing, pp. 15–
548 26. https://doi.org/10.1007/978-3-319-46116-8_2

549 Dowdle, J., Ishikawa, T., Gatzek, S., Rolinski, S., Smirnov, N., 2007. Two genes in *Arabidopsis*
550 *thaliana* encoding GDP-l-galactose phosphorylase are required for ascorbate biosynthesis and
551 seedling viability. *Plant J.* 52, 673–689. <https://doi.org/10.1111/J.1365-313X.2007.03266.X>

552 Fageria, N., Stone, L., 2006. Physical, chemical, and biological changes in the rhizosphere and
553 nutrient availability. *J. Plant Nutr.* 29, 1327–1356.
554 <https://doi.org/10.1080/01904160600767682>

555 Fedeli, R., Vannini, A., Celletti, S., Maresca, V., Munzi, S., Cruz, C., Alexandrov, D., Guarnieri,
556 M., Loppi, S., 2022. Foliar application of wood distillate boosts plant yield and nutritional
557 parameters of chickpea. *Ann. Appl. Biol.* <https://doi.org/10.1111/AAB.12794>

558 Gallie, D.R., 2013. L-Ascorbic Acid: A Multifunctional Molecule Supporting Plant Growth and
559 Development. *Scientifica (Cairo)*. 2013, 1–24. <https://doi.org/10.1155/2013/795964>

560 Gao, M., Liu, Y., Song, Z., 2019. Effects of polyethylene microplastic on the phytotoxicity of di-n-
561 butyl phthalate in lettuce (*Lactuca sativa* L. var. *ramosa* Hort). *Chemosphere* 237, 124482.

562 <https://doi.org/10.1016/J.CHEMOSPHERE.2019.124482>

563 Hartmann, N.B., Hüffer, T., Thompson, R.C., Hassellöv, M., Verschoor, A., Daugaard, A.E., Rist,
564 S., Karlsson, T., Brennholt, N., Cole, M., Herrling, M.P., Hess, M.C., Ivleva, N.P., Lusher,
565 A.L., Wagner, M., 2019. Are We Speaking the Same Language? Recommendations for a
566 Definition and Categorization Framework for Plastic Debris. *Environ. Sci. Technol.* 53, 1039–
567 1047. [https://doi.org/10.1021/ACS.EST.8B05297/ASSET/IMAGES/MEDIUM/ES-2018-](https://doi.org/10.1021/ACS.EST.8B05297/ASSET/IMAGES/MEDIUM/ES-2018-05297K_0006.GIF)
568 [05297K_0006.GIF](https://doi.org/10.1021/ACS.EST.8B05297/ASSET/IMAGES/MEDIUM/ES-2018-05297K_0006.GIF)

569 Hasanuzzaman, M., Bhuyan, M.H.M.B., Zulfiqar, F., Raza, A., Mohsin, S.M., Al Mahmud, J.,
570 Fujita, M., Fotopoulos, V., 2020. Reactive Oxygen Species and Antioxidant Defense in Plants
571 under Abiotic Stress: Revisiting the Crucial Role of a Universal Defense Regulator.
572 *Antioxidants* 2020, Vol. 9, Page 681–9, 681. <https://doi.org/10.3390/ANTIOX9080681>

573 Hazelton, P., Murphy, B., 2007. *Interpreting Soil Test Results: What Do All the Numbers Mean?*
574 CSIRO PUBLISHING 150 Oxford Street (PO Box 1139) Collingwood VIC 3066 Australia.

575 Huerta-Lwanga, E., Mendoza-Vega, J., Ribeiro, O., Gertsen, H., Peters, P., Geissen, V., 2021. Is the
576 Polylactic Acid Fiber in Green Compost a Risk for *Lumbricus terrestris* and *Triticum*
577 *aestivum*? *Polymers (Basel)*. 13, 1–10. <https://doi.org/10.3390/POLYM13050703>

578 Huerta Lwanga, E., Mendoza Vega, J., Ku Quej, V., Chi, J. de los A., Sanchez del Cid, L., Chi, C.,
579 Escalona Segura, G., Gertsen, H., Salánki, T., van der Ploeg, M., Koelmans, A.A., Geissen, V.,
580 2017. Field evidence for transfer of plastic debris along a terrestrial food chain. *Sci. Reports*
581 2017 7:1–7. <https://doi.org/10.1038/s41598-017-14588-2>

582 Ishikawa, T., Dowdle, J., Smirnoff, N., 2006. Progress in manipulating ascorbic acid biosynthesis
583 and accumulation in plants. *Physiol. Plant.* 126, 343–355. [https://doi.org/10.1111/j.1399-](https://doi.org/10.1111/j.1399-3054.2006.00640.x)
584 [3054.2006.00640.x](https://doi.org/10.1111/j.1399-3054.2006.00640.x)

585 Jagota, S.K., Dani, H.M., 1982. A new colorimetric technique for the estimation of vitamin C using
586 Folin phenol reagent. *Anal. Biochem.* 127, 178–182. [https://doi.org/10.1016/0003-](https://doi.org/10.1016/0003-2697(82)90162-2)
587 [2697\(82\)90162-2](https://doi.org/10.1016/0003-2697(82)90162-2)

588 Jiang, X.J., Liu, W., Wang, E., Zhou, T., Xin, P., 2017. Residual plastic mulch fragments effects on
589 soil physical properties and water flow behavior in the Minqin Oasis, northwestern China. *Soil*
590 *Tillage Res.* 166, 100–107. <https://doi.org/10.1016/J.STILL.2016.10.011>

591 Khalid, N., Aqeel, M., Noman, A., 2020. Microplastics could be a threat to plants in terrestrial
592 systems directly or indirectly. *Environ. Pollut.* 267, 115653.
593 <https://doi.org/10.1016/J.ENVPOL.2020.115653>

594 Lamberti, F.M., Román-Ramírez, L.A., Wood, J., 2020. Recycling of Bioplastics: Routes and
595 Benefits. *J. Polym. Environ.* 2020 2810 28, 2551–2571. [https://doi.org/10.1007/S10924-020-](https://doi.org/10.1007/S10924-020-01795-8)
596 [01795-8](https://doi.org/10.1007/S10924-020-01795-8)

597 Lebreton, L., Slat, B., Ferrari, F., Sainte-Rose, B., Aitken, J., Marthouse, R., Hajbane, S., Cunsolo,
598 S., Schwarz, A., Levivier, A., Noble, K., Debeljak, P., Maral, H., Schoeneich-Argent, R.,
599 Brambini, R., Reisser, J., 2018. Evidence that the Great Pacific Garbage Patch is rapidly
600 accumulating plastic. *Sci. Rep.* 8, 4666. <https://doi.org/10.1038/s41598-018-22939-w>

601 Li, L., Luo, Y., Li, R., Zhou, Q., Peijnenburg, W.J.G.M., Yin, N., Yang, J., Tu, C., Zhang, Y., 2020.
602 Effective uptake of submicrometre plastics by crop plants via a crack-entry mode. *Nat. Sustain.*
603 2020 311 3, 929–937. <https://doi.org/10.1038/s41893-020-0567-9>

604 Li, S., Wang, T., Guo, J., Dong, Y., Wang, Z., Gong, L., Li, X., 2021. Polystyrene microplastics
605 disturb the redox homeostasis, carbohydrate metabolism and phytohormone regulatory
606 network in barley. *J. Hazard. Mater.* 415, 125614.
607 <https://doi.org/10.1016/J.JHAZMAT.2021.125614>

608 Li, Z., Li, Q., Li, R., Zhou, J., Wang, G., 2021. The distribution and impact of polystyrene
609 nanoplastics on cucumber plants. *Environ. Sci. Pollut. Res.* 28, 16042–16053.
610 <https://doi.org/10.1007/s11356-020-11702-2>

611 Liu, H., Yang, X., Liu, G., Liang, C., Xue, S., Chen, H., Ritsema, C.J., Geissen, V., 2017. Response
612 of soil dissolved organic matter to microplastic addition in Chinese loess soil. *Chemosphere*
613 185, 907–917. <https://doi.org/10.1016/J.CHEMOSPHERE.2017.07.064>

614 Liwarska-Bizukojc, E., 2021. Effect of (bio)plastics on soil environment: A review. *Sci. Total*
615 *Environ.* 795, 148889. <https://doi.org/10.1016/J.SCITOTENV.2021.148889>

616 Machado, R., Serralheiro, R., 2017. Soil Salinity: Effect on Vegetable Crop Growth. *Management*
617 *Practices to Prevent and Mitigate Soil Salinization. Horticulturae* 3, 30.
618 <https://doi.org/10.3390/horticulturae3020030>

619 Meng, F., Yang, X., Riksen, M., Xu, M., Geissen, V., 2021. Response of common bean (*Phaseolus*
620 *vulgaris* L.) growth to soil contaminated with microplastics. *Sci. Total Environ.* 755, 142516.
621 <https://doi.org/10.1016/J.SCITOTENV.2020.142516>

622 Moore, B., Zhou, L., Rolland, F., Hall, Q., Cheng, W.H., Liu, Y.X., Hwang, I., Jones, T., Sheen, J.,
623 2003. Role of the *Arabidopsis* glucose sensor HXK1 in nutrient, light, and hormonal signaling.
624 *Science* (80-.). 300, 332–336.
625 https://doi.org/10.1126/SCIENCE.1080585/SUPPL_FILE/MOORE.SOM.PDF

626 Moshood, T.D., Nawanir, G., Mahmud, F., Mohamad, F., Ahmad, M.H., AbdulGhani, A., 2022.
627 Biodegradable plastic applications towards sustainability: A recent innovations in the green
628 product. *Clean. Eng. Technol.* 6, 100404. <https://doi.org/10.1016/J.CLET.2022.100404>

629 Mroczkowska, M., Germaine, K., Culliton, D., Duarte, T.K., Neves, A.C., 2021. Assessment of
630 biodegradation and eco-toxic properties of novel starch and gelatine blend bioplastics.
631 *Recycling* 6. <https://doi.org/10.3390/recycling6040081>

632 Muncke, J., Andersson, A.M., Backhaus, T., Boucher, J.M., Carney Almroth, B., Castillo Castillo,
633 A., Chevrier, J., Demeneix, B.A., Emmanuel, J.A., Fini, J.B., Gee, D., Geueke, B., Groh, K.,
634 Heindel, J.J., Houlihan, J., Kassotis, C.D., Kwiatkowski, C.F., Lefferts, L.Y., Maffini, M. V.,
635 Martin, O. V., Myers, J.P., Nadal, A., Nerin, C., Pelch, K.E., Fernández, S.R., Sargis, R.M.,
636 Soto, A.M., Trasande, L., Vandenberg, L.N., Wagner, M., Wu, C., Zoeller, R.T., Scheringer,
637 M., 2020. Impacts of food contact chemicals on human health: A consensus statement.
638 *Environ. Heal. A Glob. Access Sci. Source* 19, 1–12. [https://doi.org/10.1186/s12940-020-](https://doi.org/10.1186/s12940-020-0572-5)
639 [0572-5](https://doi.org/10.1186/s12940-020-0572-5)

640 Murray, J.E., Laurieri, N., Delgoda, R., 2017. Proteins. *Pharmacogn. Fundam. Appl. Strateg.* 477–
641 494. <https://doi.org/10.1016/B978-0-12-802104-0.00024-X>

642 Ng, E.L., Huerta Lwanga, E., Eldridge, S.M., Johnston, P., Hu, H.W., Geissen, V., Chen, D., 2018.
643 An overview of microplastic and nanoplastic pollution in agroecosystems. *Sci. Total Environ.*
644 627, 1377–1388. <https://doi.org/10.1016/J.SCITOTENV.2018.01.341>

645 Parida, A.K., Das, A.B., 2005. Salt tolerance and salinity effects on plants: a review. *Ecotoxicol.*
646 *Environ. Saf.* 60, 324–349. <https://doi.org/10.1016/J.ECOENV.2004.06.010>

647 Parisi, O.I., Morelli, C., Scrivano, L., Sinicropi, M.S., Cesario, M.G., Candamano, S., Puoci, F.,
648 Sisci, D., 2015. Controlled release of sunitinib in targeted cancer therapy: smart magnetically
649 responsive hydrogels as restricted access materials. *RSC Adv.* 5, 65308–65315.
650 <https://doi.org/10.1039/C5RA12229E>

651 Pignattelli, S., Broccoli, A., Piccardo, M., Terlizzi, A., Renzi, M., 2021. Effects of polyethylene
652 terephthalate (PET) microplastics and acid rain on physiology and growth of *Lepidium*
653 *sativum*. *Environ. Pollut.* 282, 116997. <https://doi.org/10.1016/J.ENVPOL.2021.116997>

654 Qi, Y., Yang, X., Pelaez, A.M., Huerta Lwanga, E., Beriot, N., Gertsen, H., Garbeva, P., Geissen,
655 V., 2018. Macro- and micro- plastics in soil-plant system: Effects of plastic mulch film
656 residues on wheat (*Triticum aestivum*) growth. *Sci. Total Environ.* 645, 1048–1056.
657 <https://doi.org/10.1016/j.scitotenv.2018.07.229>

658 Quagliata, G., Celletti, S., Coppa, E., Mimmo, T., Cesco, S., Astolfi, S., 2021. Potential Use of
659 Copper-Contaminated Soils for Hemp (*Cannabis sativa* L.) Cultivation. *Environ.* 2021, Vol. 8,
660 Page 111 8, 111. <https://doi.org/10.3390/ENVIRONMENTS8110111>

661 Rillig, M.C., Lehmann, A., de Souza Machado, A.A., Yang, G., 2019. Microplastic effects on
662 plants. *New Phytol.* 223, 1066–1070. <https://doi.org/10.1111/NPH.15794>

663 Rosenboom, J.G., Langer, R., Traverso, G., 2022. Bioplastics for a circular economy. *Nat. Rev.*
664 *Mater.* 2022 72 7, 117–137. <https://doi.org/10.1038/s41578-021-00407-8>

665 Rout, G.R., Sahoo, S., 2015. Role of iron in plant growth and metabolism. *Rev. Agric. Sci.* 3, 1–24.

666 <https://doi.org/10.7831/ras.3.1>

667 Sami, F., Yusuf, M., Faizan, M., Faraz, A., Hayat, S., 2016. Role of sugars under abiotic stress.
668 *Plant Physiol. Biochem.* 109, 54–61. <https://doi.org/10.1016/J.PLAPHY.2016.09.005>

669 Sforzini, S., Oliveri, L., Chinaglia, S., Viarengo, A., 2016. Application of biotests for the
670 determination of soil ecotoxicity after exposure to biodegradable plastics. *Front. Environ. Sci.*
671 4, 68. <https://doi.org/10.3389/FENVS.2016.00068/XML/NLM>

672 Sgherri, C., Cecconami, S., Pinzino, C., Navari-Izzo, F., Izzo, R., 2010. Levels of antioxidants and
673 nutraceuticals in basil grown in hydroponics and soil. *Food Chem.* 123, 416–422.
674 <https://doi.org/10.1016/j.foodchem.2010.04.058>

675 Shafqat, A., Al-Zaqri, N., Tahir, A., Alsahme, A., 2021. Synthesis and characterization of starch
676 based bioplastics using varying plant-based ingredients, plasticizers and natural fillers. *Saudi J.*
677 *Biol. Sci.* 28, 1739–1749. <https://doi.org/10.1016/J.SJBS.2020.12.015>

678 Sharma, P., Jha, A.B., Dubey, R.S., Pessarakli, M., 2012. Reactive Oxygen Species, Oxidative
679 Damage, and Antioxidative Defense Mechanism in Plants under Stressful Conditions. *J. Bot.*
680 2012, 1–26. <https://doi.org/10.1155/2012/217037>

681 Shulaev, V., Oliver, D.J., 2006. Metabolic and proteomic markers for oxidative stress. New tools
682 for reactive oxygen species research. *Plant Physiol.* 141, 367–372.
683 <https://doi.org/10.1104/PP.106.077925>

684 Spaccini, R., Todisco, D., Drosos, M., Nebbioso, A., Piccolo, A., 2016. Decomposition of bio-
685 degradable plastic polymer in a real on-farm composting process. *Chem. Biol. Technol. Agric.*
686 3, 1–12. <https://doi.org/10.1186/s40538-016-0053-9>

687 Su, L.J., Zhang, J.H., Gomez, H., Murugan, R., Hong, X., Xu, D., Jiang, F., Peng, Z.Y., 2019.
688 Reactive Oxygen Species-Induced Lipid Peroxidation in Apoptosis, Autophagy, and
689 Ferroptosis. *Oxid. Med. Cell. Longev.* 2019. <https://doi.org/10.1155/2019/5080843>

690 Sun, X.D., Yuan, X.Z., Jia, Y., Feng, L.J., Zhu, F.P., Dong, S.S., Liu, J., Kong, X., Tian, H., Duan,
691 J.L., Ding, Z., Wang, S.G., Xing, B., 2020. Differentially charged nanoplastics demonstrate

692 distinct accumulation in *Arabidopsis thaliana*. *Nat. Nanotechnol.* 2020 15, 755–760.
693 <https://doi.org/10.1038/s41565-020-0707-4>

694 Thompson, R.C., Moore, C.J., Vom Saal, F.S., Swan, S.H., 2009. Plastics, the environment and
695 human health: current consensus and future trends. *Philos. Trans. R. Soc. B Biol. Sci.* 364,
696 2153–2166. <https://doi.org/10.1098/rstb.2009.0053>

697 Thompson, R.C., Olson, Y., Mitchell, R.P., Davis, A., Rowland, S.J., John, A.W.G., McGonigle,
698 D., Russell, A.E., 2004. Lost at Sea: Where Is All the Plastic? *Science* (80-.). 304, 838.
699 https://doi.org/10.1126/SCIENCE.1094559/SUPPL_FILE/THOMPSON.SOM.PDF

700 Ullah, R., Tsui, M.T.K., Chen, H., Chow, A., Williams, C., Ligaba-Osen, A., 2021. Microplastics
701 interaction with terrestrial plants and their impacts on agriculture. *J. Environ. Qual.* 50, 1024–
702 1041. <https://doi.org/10.1002/jeq2.20264>

703 van Weert, S., Redondo-Hasselerharm, P.E., Diepens, N.J., Koelmans, A.A., 2019. Effects of
704 nanoplastics and microplastics on the growth of sediment-rooted macrophytes. *Sci. Total*
705 *Environ.* 654, 1040–1047. <https://doi.org/10.1016/J.SCITOTENV.2018.11.183>

706 Wang, F., Zhang, X., Zhang, Shuqi, Zhang, Shuwu, Sun, Y., 2020. Interactions of microplastics and
707 cadmium on plant growth and arbuscular mycorrhizal fungal communities in an agricultural
708 soil. *Chemosphere* 254. <https://doi.org/10.1016/J.CHEMOSPHERE.2020.126791>

709 Wang, Y., Ding, K., Ren, L., Peng, A., Zhou, S., 2022. Biodegradable Microplastics: A Review on
710 the Interaction with Pollutants and Influence to Organisms. *Bull. Environ. Contam. Toxicol.*
711 108, 1006–1012. <https://doi.org/10.1007/s00128-022-03486-7>

712 Wani, A.S., Ahmad, A., Hayat, S., Fariduddin, Q., 2013. Salt-induced modulation in growth,
713 photosynthesis and antioxidant system in two varieties of *Brassica juncea*. *Saudi J. Biol. Sci.*
714 20, 183–193. <https://doi.org/10.1016/J.SJBS.2013.01.006>

715 Wu, X., Liu, Y., Yin, S., Xiao, K., Xiong, Q., Bian, S., Liang, S., Hou, H., Hu, J., Yang, J., 2020.
716 Metabolomics revealing the response of rice (*Oryza sativa* L.) exposed to polystyrene
717 microplastics. *Environ. Pollut.* 266, 115159. <https://doi.org/10.1016/J.ENVPOL.2020.115159>

718 Yu, H., Fan, P., Hou, J., Dang, Q., Cui, D., Xi, B., Tan, W., 2020. Inhibitory effect of microplastics
719 on soil extracellular enzymatic activities by changing soil properties and direct adsorption: An
720 investigation at the aggregate-fraction level. *Environ. Pollut.* 267, 115544.
721 <https://doi.org/10.1016/J.ENVPOL.2020.115544>

722 Zhang, D., Liu, H. bin, Hu, W. li, Qin, X. hui, Ma, X. wang, Yan, C. rong, Wang, H. yuan, 2016.
723 The status and distribution characteristics of residual mulching film in Xinjiang, China. *J.*
724 *Integr. Agric.* 15, 2639–2646. [https://doi.org/10.1016/S2095-3119\(15\)61240-0](https://doi.org/10.1016/S2095-3119(15)61240-0)

725 Zimmermann, L., Dombrowski, A., Völker, C., Wagner, M., 2020. Are bioplastics and plant-based
726 materials safer than conventional plastics? In vitro toxicity and chemical composition.
727 *Environ. Int.* 145. <https://doi.org/10.1016/J.ENVINT.2020.106066>

728

729 **Figure captions**

730 **Fig. 1.** Image showing bioplastic before drying obtained by the casting technique.

731

732 **Fig. 2.** Content of total chlorophyll (A) and biometric parameters [plant height (B), leaf number (C),
733 and shoot fresh weight (D)] of basil plants grown for 35 days without (C = control, grey bar) and
734 with (B = bioplastic, green bar) corn starch-based bioplastic added to the soil at the concentration of
735 2.5% (w/w). Image (E) comparing basil plants grown without (C = control, on the left) and with
736 bioplastic (B = bioplastic, on the right) added in the soil. All data are reported as mean values \pm SE.
737 The statistical significance between C and B conditions was tested by Student's t-test ($* = p < 0.05$;
738 $*** = p < 0.001$). t-value and p-value are indicated.

739

740 **Fig. 3.** Content of soluble sugars (*i.e.*, fructose, glucose, sucrose, and their sum) (A) and total
741 soluble proteins (B) in shoots of basil plants grown for 35 days without (C = control, grey bar) and
742 with (B = bioplastic, green bar) corn starch-based bioplastic added to the soil at the concentration of
743 2.5% (w/w). All data are reported as mean values \pm SE. The statistical significance between C and
744 B conditions was tested by Student's t-test ($* = p < 0.05$; $** = p < 0.01$; ns = not significant). t-
745 value and p-value are indicated.

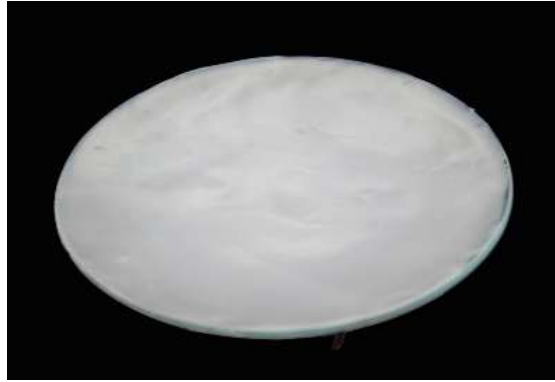
746

747 **Fig. 4.** Content of ascorbic acid (vitamin C) (A) and malondialdehyde (MDA) (B) in shoots of basil
748 plants grown for 35 days without (C = control, grey bar) and with (B = bioplastic, green bar) corn
749 starch-based bioplastic added to the soil at the concentration of 2.5% (w/w). All data are reported as
750 mean values \pm SE. The statistical significance between C and B conditions was tested by Student's
751 t-test ($* = p < 0.05$). t-value and p-value are indicated. Correlation analysis between the shoot
752 content of ascorbic acid and MDA (C) of basil plants grown for 35 days without (C = control, grey
753 dot) and with (B = bioplastic, green dot) corn starch-based bioplastic added to the soil at the
754 concentration of 2.5% (w/w). Pearson correlation coefficient (r) and p-value are indicated.

755

756

757 **Fig. 5.** pH (A), electrical conductivity – EC (B), and cation exchange capacity – CEC (C) of the soil
758 where basil plants were grown for 35 days without (C = control, grey bar) and with (B = bioplastic,
759 green bar) corn starch-based bioplastic at the concentration of 2.5% (w/w). All data are reported as
760 mean values \pm SE. The statistical significance between C and B conditions was tested by Student's
761 t-test (* = $p < 0.05$; *** = $p < 0.001$; ns = not significant). t-value and p-value are indicated.

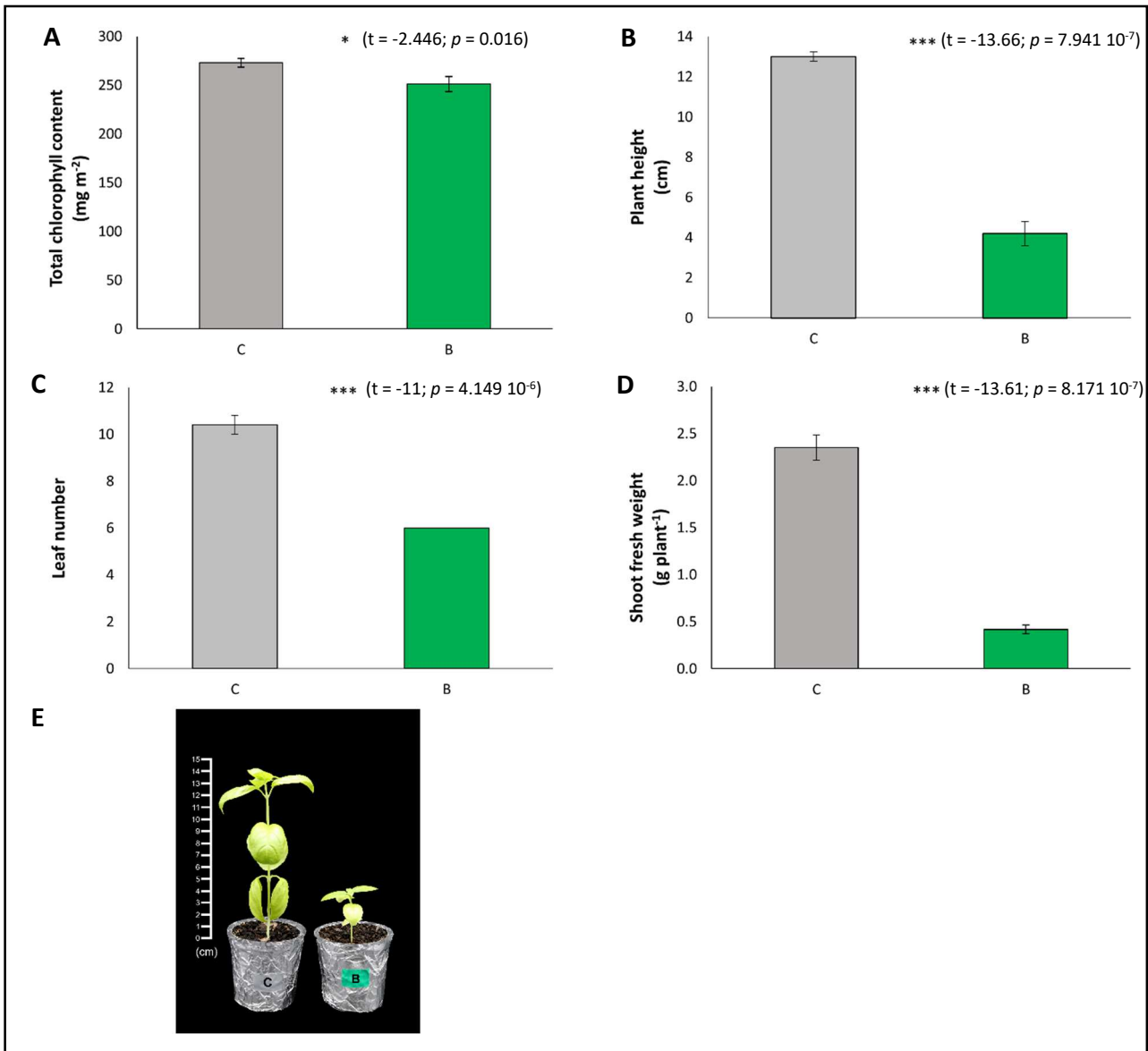


762

763 **Fig. 1**

764

765

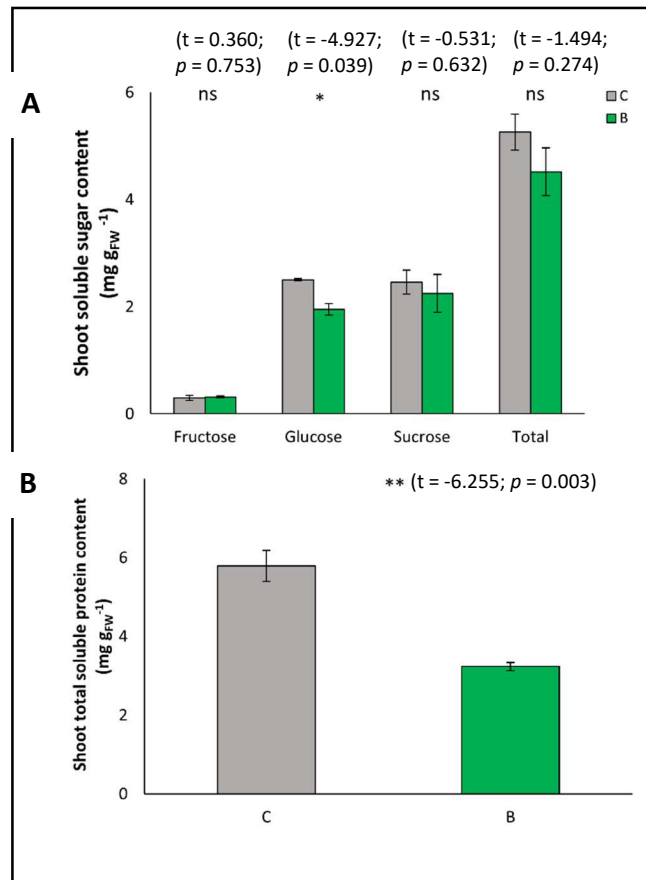


766

767 **Fig. 2**

768

769



770

771 **Fig. 3**

772

773

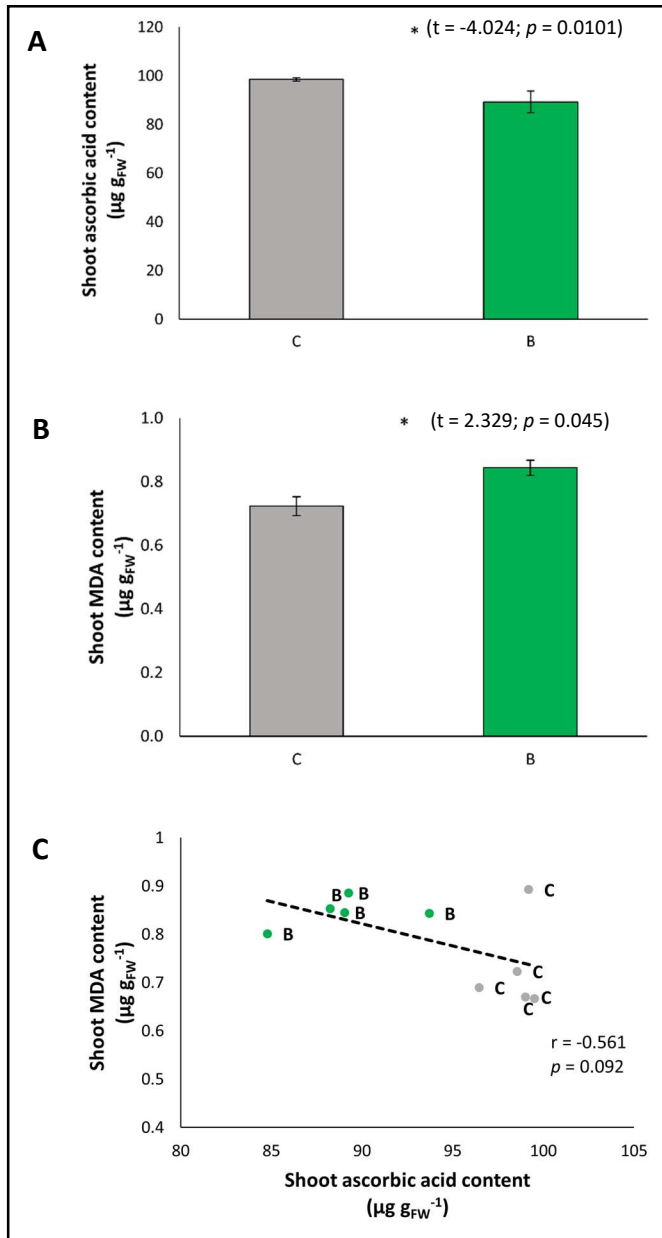
774

775

776

777

778



779 **Fig. 4**

780

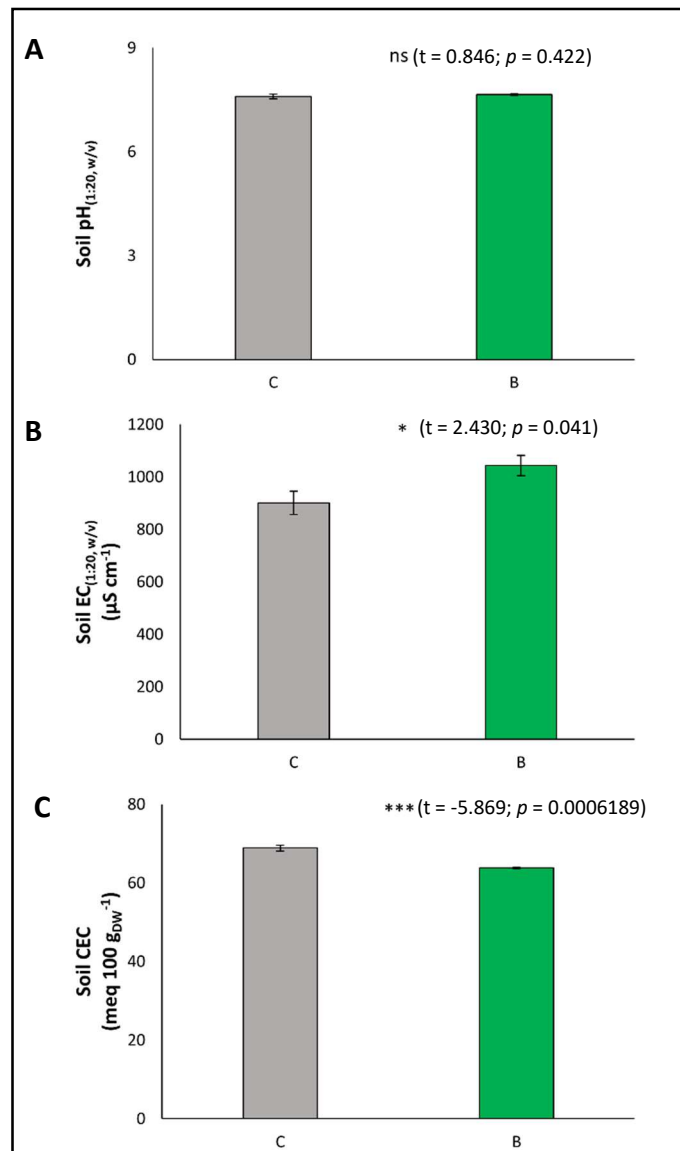
781

782

783

784

785



786

787 **Fig. 5**